1	Evidence of host switching: sea lampreys disproportionately attack
2	Chinook salmon when lake trout abundance is low in Lake Ontario
3	Draft: January 18, 2020
4	Jean V. Adams ^a *, Michael L. Jones ^b
5	^a US Geological Survey - Great Lakes Science Center, 1451 Green Road, Ann Arbor, MI
6	48105, USA, jvadams@usgs.gov, 734-994-3331
7	^b Michigan State University, 13 Natural Resources Building, East Lansing, MI 48824, USA
8	Disclaimer
9	This draft manuscript is distributed solely for purposes of scientific peer review. Its content is

- 10 *deliberative and predecisional, so it must not be disclosed or released by reviewers. Because*
- 11 the manuscript has not yet been approved for publication by the U.S. Geological Survey
- 12 *(USGS), it does not represent any official USGS finding or policy.*

13 Abstract

- 14 Lake trout (Salvelinus namaycush) is the presumed preferred host of the invasive sea
- 15 lamprey (*Petromyzon marinus*) in the Great Lakes, but little is understood about this
- 16 preference outside of laboratory experiments. By preference we mean sea lamprey attacks
- 17 on hosts are disproportionate to host relative abundance. The purpose of this study was to
- 18 quantify host preference of sea lampreys in the field for the first time. We focused our

19 analysis on Lake Ontario, where the two dominant host species for sea lampreys are lake 20 trout and Chinook salmon (Oncorhynchus tshawytscha). Sea lampreys exhibited a strong 21 preference for lake trout when host abundance was $\geq 32\%$ lake trout, but sea lamprey 22 preference switched to Chinook salmon when host abundance was $\leq 13\%$ lake trout. 23 Model results were equivocal with respect to determining whether the primary predictor 24 of sea lamprey preference was relative or absolute abundance of lake trout. Other evidence, 25 particularly the different spatial distributions of the two hosts, suggests that sea lamprey 26 preference for lake trout is based on a higher encounter probability, driven by absolute 27 abundance of lake trout. This study confirms a widely held suspicion that observed sea 28 lamprey-induced marking rates on lake trout, used to assess the status of sea lamprey 29 control in the Great Lakes, can be influenced by the abundance of alternative hosts. As an 30 alternative host for sea lamprey parasitism, Chinook salmon may contribute to the 31 persistence of lake trout in Lake Ontario.

32 Keywords

33 predation; parasitism; prey preference; prey switching; Great Lakes

34 Introduction

Lake trout (*Salvelinus namaycush*) are presumed to be the preferred hosts for invasive sea
lampreys (*Petromyzon marinus*) in the Great Lakes (Harvey et al., 2008; Pycha and King,
1975; Swink, 2003). But, juvenile sea lampreys attack many other Great Lakes species,

especially large-bodied ones (Farmer and Beamish, 1973; Swink, 2003), and population
level impacts have been observed or hypothesized for several species: Chinook salmon
(*Oncorhynchus tshawytscha*, McLeod et al., 2011), lake whitefish (*Coregonus clupeaformis*,
Spangler et al., 1980; Ebener et al., 2010), burbot (*Lota lota*, Stapanian et al., 2006), and
coregonines and suckers (*Coregonus* spp. and *Catostomus* spp., Harvey et al., 2008).

43 In spite of their pelagic habitat and fast swimming speeds, Chinook salmon are commonly attacked by sea lampreys. Most (62%) of the age-2 precocious Chinook salmon had sea 44 45 lamprey-induced marks in Lake Huron in 1970 (Smith and Tibbles, 1980). Sea lamprey 46 marking rates on Lake Huron Chinook salmon ranged from 0.14 to 0.33 marks per fish in 47 1988-1992 (Johnson et al., 1995). The abundance of juvenile sea lampreys was significantly 48 positively correlated with Chinook salmon abundance in Lake Huron, suggesting that 49 Chinook salmon could be an important component of the sea lamprey diet (Young et al., 50 1996).

51 Little is understood about how sea lamprey parasitism varies with host abundance (Bence 52 et al., 2003; Koonce, 1987). Knowledge in this area would allow us to improve our 53 estimates of host damage. Current estimates based on observed sea lamprey marking rates 54 are subject to survivor bias (marks are only observed on hosts that survived attacks, 55 Adams et al., *This issue*) and from sampling bias (few host species other than lake trout are 56 regularly surveyed on a lake-wide spatial scale). Although the focus of this study is 57 predator-prey dynamics, we refer to sea lampreys as parasites (rather than predators) and 58 lake trout and Chinook salmon as hosts (rather than prev).

59 The purpose of this study was to take the first step toward quantifying host preference of 60 sea lamprevs in the field. In this paper, the term preference is used to indicate sea lamprev 61 attacks on hosts disproportionate to host relative abundance. In order to see a switch in 62 preference in the field, feeding observations must be made over a wide range of relative 63 abundances of hosts. That's why so many switching studies are carried out in the 64 laboratory where the experimenter can control prey densities (Bayliss, 1982). We chose 65 Lake Ontario as our case study, focusing on two host species: lake trout and Chinook 66 salmon. Lake Ontario has a history of a relatively stable sea lamprey population, with 67 variable lake trout abundance and increasing Chinook salmon abundance (Lantry et al., 68 2015). These conditions present a prime opportunity to observe host switching in sea 69 lampreys, wherein sea lampreys switch their preference from lake trout to Chinook salmon 70 below some threshold of the relative abundance of lake trout.

Our objectives were to characterize the preference of sea lampreys for lake trout in Lake Ontario and determine if and when their preference switches to Chinook salmon based on field observations. Greater understanding of host preference may aid in efforts to control sea lampreys and manage their hosts in the Great Lakes (Stewart et al., 2003) as well as to inform conservation of sea lampreys in their native range (Maitland et al., 2015).

76 Methods

77 Evidence of host preference and host switching was explored using logistic regression

78 models of sea lamprey parasitism on lake trout and Chinook salmon in Lake Ontario. Recall

that we defined preference as sea lamprey attacks disproportionate to host relative abundance. The two key quantities are the proportion of hosts that are lake trout and the proportion of sea lamprey-induced marks that are on lake trout, assuming that lake trout and Chinook salmon are the only available hosts. The proportion of marks on lake trout was assumed to be a function of the proportion of hosts that were lake trout,

(1)
$$ln\left(\frac{M_1}{1-M_1}\right) = \gamma_0 + \gamma_1 ln\left(\frac{P_1}{1-P_1}\right)$$

84 where M_1 is the proportion of sea lamprey-induced marks on lake trout,

$$M_1 = \frac{S_1 O_1 / E_1}{S_1 O_1 / E_1 + S_2 O_2 / E_2} ,$$

 S_i is the estimated lake-wide abundance of species *i* (*i* = 1 for lake trout; *i* = 2 for Chinook 85 salmon), O_i is the observed number of marks on E_i examined individuals of species *i*, 86 87 $P_1 = S_1 / \sum S_i$ is the proportion of host abundance that were lake trout, and γ_0 and γ_1 are 88 parameters to be estimated. When fitting the logistic regression, the response was scaled to 89 the total number of observed marks, $\sum O_i$, as the sample size. Because the scaled responses 90 were not integers, the quasibinomial family was used in the general linear model rather 91 than the usual binomial family. This is a natural log transformation of the simple ratio predation model (Murdoch, 1969) with the addition of a switching parameter, γ_1 . If $\gamma_1 \neq 1$, 92 93 there is evidence of a switch in host preference that depends on the proportion of hosts that were lake trout. If $\gamma_1 = 1$, there is no host switching, and preference over all host 94 proportions is simply estimated by e^{γ_0} (equivalent to the constant c in the simple ratio 95 predation model, Murdoch, 1969). If $\gamma_1 = 1$ and $\gamma_0 > 0$, then $M_1 > P_1$ and sea lampreys 96

97 exhibit a preference for lake trout; if $\gamma_1 = 1$ and $\gamma_0 < 0$, then $M_1 < P_1$ and they exhibit a 98 preference for Chinook salmon. Scatter plots of residuals from the full model were visually 99 examined for qualitative evidence of patterns related to year, lake trout abundance, 100 Chinook salmon abundance, adult sea lamprey relative abundance, and adult sea lamprey 101 size.

102 Lake-wide abundance of hosts was estimated by two separate Lake Ontario statistical 103 catch-at-age models (Table 1), one for lake trout (Brenden et al., 2011, Travis Brenden, 104 Michigan State University, personal communication), and one for Chinook salmon 105 (Kimberly Fitzpatrick, Cornell University, personal communication; estimates are 106 preliminary and have not been endorsed by the Lake Ontario Committee or the Lake 107 Ontario Technical Committee; data can be found in Connerton, 2019; Bishop et al., 2019; 108 Connerton and Eckert, 2019; Lake, 2017; Prindle and Bishop, 2019; Sanderson et al., 2019; 109 Yuille and Holden, 2017). Abundance of lake trout was calculated as the sum of age 3 and 110 older fish (Brenden et al., 2011). Abundance of Chinook salmon was calculated as the sum 111 of age 1 and older fish, corresponding to the spawning adults which were surveyed. For 112 both species, the fish included in the abundance estimates are susceptible to sea lamprey 113 predation.

Host parasitism rates were derived from observed sea lamprey marks on hosts. Host
parasitism was estimated as the number of A1 marks per host > 431 mm, following current
practice for Lake Ontario (Lantry et al., 2015; Larson et al., 2003). This rate is calculated as
the total number of marks (including multiple marks on a single host) divided by the total

118 number of fish examined for marks. The A1 classification indicates that the marks pierced 119 the skin and are "fresh" or little healed (Ebener et al., 2003; King Jr., 1980). For lake trout, 120 observations of marks (Table 1) and measures of total length (Figure 1) came from two 121 fishery independent gill net surveys: the Lake Ontario and Bay of Quinte fish community 122 index gill netting survey conducted during June-September in Canada waters (Hoyle, 2018) 123 and the lake trout assessment survey conducted during September-October in US waters 124 (Lantry et al., 2018). For Chinook salmon, observations of marks and measures of total 125 length came from the fall spawning migration up a fish ladder on Beaverdam Brook (a 126 tributary of the Salmon River) into the New York State Department of Environmental 127 Conservation Salmon River Fish Hatchery, Albion NY, (Prindle and Bishop, 2018). 128 The relative abundance of adult sea lampreys in Lake Ontario was assessed annually by 129 mark-recapture studies in five index streams (Humber River, Duffins Creek, and 130 Bowmanville River in Canada and Black and Sterling Rivers in the US) during their 131 spawning migration (Mullett et al., 2003). The spawning migration occurs after sea 132 lampreys spend 6-18 months in the lake feeding on fish as juveniles. The stream pooled 133 Petersen estimates (Seber, 1970) were then summed for a lake-wide adult index (Barber et 134 al., This issue). Individual sea lampreys captured during these studies were weighed, and 135 the median annual weight was used as an index of sea lamprey size in the exploration of 136 model residuals.

Additional models were fit to the parasite-host data that accounted for the sizes of the two
host species, because Great Lakes sea lampreys have a demonstrated preference for large

fish (Bence et al., 2003; Farmer and Beamish, 1973; Rutter and Bence, 2003). Without
accounting for host size, we expect that sea lamprey preference for lake trout will be
underestimated, because Chinook salmon tend to be larger than lake trout in Lake Ontario
(Figure 1).

Expected marking rates were modeled as a logistic function of host size (Rutter and Bence,
2003) for both host species combined (allowing marking rates to vary with host size
regardless of species) and for each host species separately (allowing for species-specific
size relations),

$$W = \frac{\theta}{1 + exp[-\alpha(L-\beta)]},$$

where *W* is the expected number of marks following a negative binomial distribution
(Prichard and Bence, 2013), *L* is host length (in mm), *θ* is the asymptotic marking rate, *β* is
the length of the host at the inflection point, and *αθ*/4 is the slope at the inflection point.
Models were fit using maximum likelihood with the nlminb function of R (R Core Team,
2018). All three parameters were estimated in the natural log-transformed space to
constrain them to be positive.

153 The expected proportion of sea lamprey-induced marks on lake trout based on the154 abundance and size distribution of the host species was calculated as

$$\tilde{P}_{1, \ size} = \frac{S_1 W_{13}}{S_1 \widetilde{W}_{13} + S_2 \widetilde{W}_{23}}$$
 ,

$$\tilde{P}_{1, \ size + species} = \frac{S_1 \tilde{W}_{11}}{S_1 \tilde{W}_{11} + S_2 \tilde{W}_{22}},$$

155 where \widetilde{W}_{ig} is the annual mean predicted marking rate for species *i* calculated for group *g* of

156 the data (g = 1 for lake trout alone, g = 2 for Chinook salmon alone, and g = 3 for both

157 species combined). We then fit two additional parasitism models by replacing the P_1 in

158 Model 1 with the expected proportions from asymptotic marking rates depending on host

size alone (Model 2, $\tilde{P}_{1, size}$) and on host size and species (Model 3, $\tilde{P}_{1, size + species}$).

160 Finally, we fit two additional logistic regression models, where the independent variable of

161 the relative abundance of lake trout, $ln[P_1/(1 - P_1)]$, was replaced by the absolute

abundance of lake trout ($ln(S_1)$, Model 4) and by the absolute abundance of Chinook

163 salmon ($ln(S_2)$, Model 5).

164 **Results**

165 Without taking host size into account (Model 1), the estimate for γ_1 was significantly 166 different from 1 (5.54 with standard error 1.37), indicating a switch in host preference. Sea 167 lampreys exhibited a preference for Chinook salmon when the combined abundance was \leq 168 13% lake trout, switching to a preference for lake trout when combined abundance was \geq 169 32% lake trout (γ_0 = 4.95 with standard error 1.30, Figure 2). The steepest increase in the 170 proportion of marks on lake trout occurred when the combined abundance was 32.6% lake 171 trout. No linear or nonlinear relations were evident in plots of the residuals from the full model versus year, lake trout abundance, Chinook salmon abundance, sea lampreyabundance, or sea lamprey size.

174 Sea lamprey preference for lake trout was also evident when modeling marking rates as a 175 function of host size (Figure 3). Marking rates on lake trout most steeply increased at a host 176 length 132 mm shorter than that on Chinook salmon (Table 2). The asymptotic marking 177 rate of lake trout was 1.4 times that of Chinook salmon. However, taking host size into 178 account, either for both species together (Model 2, residual deviance 48.8) or each species 179 separately (Model 3, 50.1) did not improve the fit achieved with Model 1 (residual deviance 180 47.4, t = 3.82, P = 0.0021, Figure 4), even with the inclusion of additional estimated 181 parameters.

182 The proportion of marks on lake trout was also well explained by the absolute abundance 183 of lake trout, with slightly less residual deviance (Model 4, residual deviance 47.3, t = 3.86, 184 P = 0.0020, Figure 5) than by the proportional abundance of lake trout (Model 1). The 185 estimate for γ_1 was 1.78 with standard error 0.43. Here, the γ_1 parameter loses its 186 interpretation as a switching indicator, because this model depends on the abundance of 187 only one host species. The estimate for γ_0 was 4.73 with standard error 1.22. The steepest 188 increase in the proportion of marks on lake trout occurred when the absolute abundance of 189 lake trout was 0.69 million fish.

190 In contrast, the absolute abundance of Chinook salmon was not a strong predictor of the 191 proportion of marks on lake trout (Model 5, residual deviance 85.7, t = -1.76, P = 0.10).

192 **Discussion**

193 This study provides an example of a parasite with a strong host preference that switches 194 when the relative and absolute abundance of the preferred prey is low. This is in contrast 195 to the prediction by Murdoch (1969) that prey switching will not occur when the predator 196 has strong prey preferences (Murdoch et al., 1975; Rubega and Inouye, 1994). In spite of 197 their strong host preference in the Great Lakes, we consider the sea lamprey to be a 198 generalist predator, attacking a wide range of fish species (Schoener, 1971). This varied 199 diet is supported by studies of stable isotopes in Lake Superior (Harvey et al., 2008) and 200 fatty acids in Lake Michigan (Happel et al., 2017). In the Atlantic, sea lamprey attacks have 201 been documented on 54 host species (Silva et al., 2014). When prey abundance is high, 202 most generalist predators feed on a single primary prey (Andersson and Erlinge, 1977). If 203 that primary prey population declines, predators turn to less abundant alternatives 204 (Angelstam et al., 1984). Switching behavior has been documented for several so-called 205 generalist predators (Cornell, 1976; Murdoch, 1969; Patterson et al., 1998). Anecdotal 206 evidence of host switching was observed in South Bay of Lake Huron when sea lamprey-207 induced marks on white suckers (*Catostomus commersonii*) coincided with the crash of the 208 lake trout population in the 1950s (Coble, 1967).

Evidence for host-switching is based on the fit of the observed data to a switching model.
For Lake Ontario during 2000-2014, Model 1 predicted that sea lampreys preferred lake
trout in 11 of 15 years and exhibited no preference between lake trout and Chinook salmon
the other four years (2006-2008 and 2014). None of the 15 years had a low enough

213 proportion of hosts that were lake trout for Model 1 to predict sea lamprey preference for 214 Chinook salmon. Thus, inferring that sea lampreys prefer attacking Chinook salmon if lake 215 trout are relatively scarce is based on extrapolation beyond the range of data used to build 216 the model. However, we did observe more marks on Chinook salmon than expected from 217 their relative abundance in three years (2007-2009, the three points below the dotted line 218 in Figure 2). These years likely correspond to more relatively small fish in the lake trout 219 population, contributing to the rise in lake trout abundance after 2009 (Brian Lantry, US 220 Geological Survey, personal communication.)

Switching to an alternative prey species makes sense for a number of reasons, most of
which lead to an increase in feeding efficiency (Bayliss, 1982). According to optimal
foraging theory, a predator chooses the prey that will give the maximum net benefit to the
individual (Begg et al., 2003). The net benefit depends on handling time, nutritional value,
and the abundance of alternative prey (Van Baalen et al., 2001).

Handling time incorporates the pursuit, capture, and digestion of prey (Fujii et al., 1986).
Chinook salmon are much more active than lake trout, searching open water habitat for
active pelagic prey (Stewart and Ibarra, 1991), which may make them more difficult for sea
lampreys to pursue and attack. Although we found no laboratory studies of sea lampreys
feeding on Chinook salmon, relatively low detachment rates (< 18%) have been observed
for several species except for burbot (36%) and walleye (*Sander vitreus*, 75%) (Farmer and
Beamish, 1973).

233 Switching to the most common prey species may be a consequence of the predator's 234 searching behavior, e.g., the use of a search image of the locally most abundant prey species 235 (May, 1977). A decrease in prey biomass results in an increase in the foraging costs of the 236 preferred prey through increased search time (Krebs and Davies, 1987). Predators may 237 forage in different patches, form prey search images, decrease prey handling time with 238 experience, or increase searching efficiency within patches (Joern, 1988). Predators are 239 more likely to detect prey if the prey are aggregated (Taylor, 1977). When prey are 240 clumped, nearly double the number of predators have been observed switching to the 241 alternative prey (Bayliss, 1982). If sea lampreys learn, their attack success rate might 242 increase with successive attempts on a given species, as has been observed in the water 243 boatman (Notonecta glauca, Hughes, 1979). A predator may also be more efficient at 244 locating prey, if it concentrates on only one species (Bayliss, 1982).

245 Switching may also be the consequence of differences in energetic content of the prey. If 246 the value of two prey types are quite different, we would expect the predator to specialize 247 on the most valuable prev type and generalize when the latter is scarce (Hughes, 1979; 248 Murdoch et al., 1975). The lipid content of lake trout was greater than that of Chinook 249 salmon in Lake Ontario in 2013 (Mumby et al., 2018) and in Canadian waters of Lake 250 Ontario during 1978-2008 (Neff et al., 2012). This higher energy content comes at the price 251 of lower metabolic and digestion rates (Fall and Fiksen, 2020) for sea lampreys feeding at 252 the lower water temperatures lake trout inhabit (to be discussed later). If lake trout were 253 more profitable prey than Chinook salmon, then we might expect to see an effect on the 254 size of the adult sea lampreys. In this case, the median weight of adult sea lampreys

collected in Lake Ontario was not significantly correlated with the proportion of marks on lake trout (r=0.34, df=13, P=0.21). This supports the contention that prey preference is not necessarily reflected in comparative growth rates (Bayliss, 1982).

258 Building a switching model that depends on the relative abundance of the prey implies that 259 the predator perceives the relative abundance of prev. The ability to do this depends on the 260 extent to which the predator and prev species overlap spatially. The diversity of hosts 261 attacked will reflect both their relative abundance and spatial distribution (Schoener, 262 1971). Sea lampreys captured in the open waters of Lake Superior fed primarily on lake 263 trout, but those in Black Bay relied heavily on coregonines and suckers (Harvey et al., 264 2008). Laboratory preference studies focus on the active choice of predators when faced 265 with more than one vulnerable prey species (e.g., Farmer and Beamish, 1973). 266 Vulnerability of different species to sea lamprey attack in the field is largely unknown. If 267 host populations were partitioned into vulnerable and invulnerable components, foraging 268 arena theory asserts that the attack rates then depend on the exchange rates between the 269 components (Ahrens et al., 2012). If recently metamorphosed sea lamprevs (transformers) 270 attack the first host fish they encounter upon exiting their natal stream, then host 271 preference is less an active choice, and more the result of passive encounter probabilities 272 and habitat overlap. For example, transformers are commonly observed attached to bloater 273 (Coregonus hoyi) captured in bottom trawling in the Detour Channel of northern Lake 274 Huron (Fleischer, 1993). Farmer and Beamish (1973) contended that sea lamprey 275 preference in the Great Lakes depended largely on the ecological distribution of host 276 species. Weitkamp et al. (2015) asserted that marking rates by western river (*Lampetra*

ayresii) and Pacific (*Entosphenus tridentatus*) lampreys in the Columbia River estuary were
influenced by the habitat selection of potential host species.

279 Lake trout and Chinook salmon occupy different habitats in Lake Ontario. Lake trout 280 occupy the hypolimnion, near the bottom of the thermocline, and Chinook salmon occupy 281 the metalimnion to maximize their growth rate potential in Lake Ontario (Mason et al., 282 1995). Chinook salmon feed almost exclusively on alewife in offshore pelagic areas; lake 283 trout consume a broader mix of alewife (*Alosa pseudoharengus*) and other species, feeding 284 on both pelagic and benthic prey in the hypolimnion (Mumby et al., 2018; Rand and 285 Stewart, 1998). The fatty acid profiles of the two host species reflect this, indicating lake 286 trout feed more on rainbow smelt (Osmerus mordax) and round gobies (Neogobius 287 *melanostomus*) than Chinook salmon, which feed more on alewives (Happel et al., 2016). 288 The energy content of the host species likely changed during 2000-2104 due to changes in 289 the composition of their prey: alewife condition increased, rainbow smelt abundance 290 declined, and round goby biomass increased in Lake Ontario during 2000-2014 (Weidel et 291 al., 2018). Sea lampreys are poorer swimmers than most teleosts and do not possess swim 292 bladders, so they must expend energy just to maintain position in the water column 293 (Beamish, 1974). Thus, it's energetically advantageous for free swimming juvenile sea 294 lampreys to search for hosts near the lake bottom, where they are less likely to encounter 295 Chinook salmon. However, sea lampreys that feed exclusively on hosts in the hypolimnnion 296 have lower growth rates than those feeding on hosts in the thermocline as a result of 297 thermal bioenergetics (Kitchell and Breck, 1980).

298 This lends theoretical support to the model based on lake trout abundance alone, 299 regardless of Chinook salmon abundance (Model 4). Similar outcomes have been observed 300 in other species. For example, the mollusk (*Lepsiella vinosa*) fed heavily on its preferred 301 species, irrespective of the relative abundance of the alternative species (Bayliss, 1982), 302 and coyotes (Canis latrans) only switched to an alternate prey when the absolute 303 abundance of its preferred prey was low (Patterson et al., 1998). Merilaita and Ruxton 304 (2009) suggested frequency-dependent predation should be common where predators face 305 large temporal or spatial fluctuations of prev types. Further evidence from the field would 306 require years of exceptionally high or low host abundance, where high absolute lake trout 307 abundance could coincide with low relative lake trout abundance or vice versa. Otherwise 308 choosing between these two models (Models 1 and 4) based on field observations alone is 309 difficult, because the relative and absolute lake trout abundances were so highly correlated 310 during 2000-2014 (r = 0.78, P = 0.0006, df = 13). If sea lampreys shift their distribution 311 when lake trout become scarce, we should be able to observe this with acoustic telemetry 312 (e.g., Holbrook et al., 2016) or archival tags (e.g., Bergstedt et al., 2012).

Sea lampreys spawning in Portuguese tributaries to the Atlantic Ocean use one of two
different feeding strategies (Lança et al., 2013). Sea lampreys either tended to attack
pelagic plankton feeders or benthic opportunist fish species. Lança et al. (2013)
hypothesized that the feeding strategies were associated with different dispersion tactics.
Sea lampreys that parasitize pelagic rather than demersal hosts enhance their dispersion
but increase the risk of not returning to fresh water. The risk of not returning to a stream to
spawn is presumably less in the more confined Great Lakes than in the Atlantic Ocean.

320 The presence of Chinook salmon as an alternative host for sea lampreys may serve as a 321 buffer, benefiting both sea lampreys and lake trout when lake trout abundance is low. By 322 concentrating attacks on the most common host, sea lampreys enable the rarer hosts to 323 enjoy higher fitness via frequency-dependent natural selection (May, 1977). Thus, the 324 presence of Chinook salmon in Lake Ontario may contribute to the persistence of lake trout 325 in the system. A predator's switch to alternative prey relieves predation pressure when the 326 prev densities are low, regardless of whether the alternative prev overlaps with the 327 primary prey (Van Baalen et al., 2001). This persistence only arises if the predator switches 328 between prey. However the rapid increase in sea lamprey attacks on lake trout at 329 abundance levels just above the switching point is an impediment to lake trout 330 reproduction. Lake trout may then be caught in a "predator pit" (Walters, 1986) and 331 maintained at low levels by sea lamprey parasitism in spite of lake trout restoration efforts. 332 Even low abundances of actively switching sea lampreys could maintain the lake trout at 333 depressed levels.

334 Lake trout and Chinook salmon are not the only species attacked by sea lampreys in Lake 335 Ontario. Boating anglers in US waters of Lake Ontario have also observed sea lampreys 336 attached to brown trout (Salmo trutta), rainbow trout (Oncorhynchus mykiss), coho salmon 337 (Oncorhynchus kisutch), and Atlantic salmon (Salmo salar) (in order of prevalence, Lantry 338 et al., 2015; Lantry and Eckert, 2018). Sea lamprey marks have also been infrequently 339 identified on other species collected in fishery-independent gill net surveys: northern pike 340 (Esox lucius), walleye, burbot (Lota lota), and lake whitefish (Hoyle, 2018; Lantry et al., 341 2018). Thus the proportions of hosts and marks that we calculate from lake trout and

Chinook salmon alone were biased low and introduced noise to our model, because the
abundances of these other hosts and the prevalence of sea lamprey-induced marks on them
has changed over time (Lantry et al., 2015). The host-switching that we detected had to
have been strong enough to overcome this background noise. The preference of Great
Lakes sea lampreys for lake trout surely contributed to our ability to detect the hostswitching, because prey switching is less pronounced when no single species dominates
the predator's diet (Hall-Scharf and Stallings, 2014; Van Leeuwen et al., 2013).

349 Different strains of lake trout in Lake Ontario were not distinguished in our analyses. The 350 Seneca strain of lake trout (from Seneca Lake, NY) has dominated lake trout stocking in 351 Lake Ontario since 1997 (Brenden et al., 2011), and has been recommended to be the 352 majority of stocked lake trout in the future (Lantry et al., 2014). Most (55%) of the lake 353 trout observed for sea lamprey marks in this study were not identified to strain. The most 354 commonly recorded strain was Seneca (20% of the total, 44% of those identified to strain), 355 with Superior strain a distant second (7% of the total, 15% of those identified to strain). 356 Seneca strain lake trout tend to be caught in deeper colder water than other strains and 357 have a lower chance of being attacked by sea lampreys and a higher chance of surviving if 358 attacked (Lantry et al., 2015; Schneider et al., 1996). However, Senecas may be less adapted 359 to conditions in Lake Ontario than previously thought (Brenden et al., 2011).

Using the observed sea lamprey marking rate as a measure of the true sea lamprey attack
rate has a number of shortcomings, but the A1 marks (indicating recent piercing attacks)
are believed to be the least affected by them. The classification of sea lamprey marks is

363 subject to error, but the A1 marks are correctly classified most of the time (Ebener et al., 364 2003, their Table 1 with four A1 marks identified correctly by 67% of 30 people on 365 average). However, the classification error associated with A1 marks exceeds that of A1-A3 366 marks combined, where the A2 and A3 marks are those that are partially healed (Adams 367 and Jones, In preparation; Firkus et al., This issue; Nowicki, 2008; Ebener et al., 2003). The 368 observed marking rate is only a good index of the true attack rate if the annual lethality 369 rate from sea lamprey attacks is relatively constant over time (Adams et al., *This issue*). 370 Fresh piercing marks (A1) should be less susceptible to survivor bias, because they may be 371 observed on fatally wounded lake trout before the lake trout die. Survivor bias may be 372 greater for Chinook salmon than lake trout, leading to underestimated marking rates, 373 because host mortality increases with water temperature (Farmer et al., 1977). The A1 374 marking rate has also been shown to be correlated with recovery rates of dead lake trout in 375 Lake Ontario (Bergstedt and Schneider, 1988). Finally, the use of A1 marks should also help 376 reduce the age-based bias in the healing rates, caused by the contrast between the fast 377 growing, young Chinook salmon and the slow growing, old lake trout in Lake Ontario. An additional complication when analyzing sea lamprey-induced marks on more than one host 378 379 species is that the duration of the juvenile (parasitic) life stage of the sea lamprey may vary 380 with host selection (Cline et al., 2014; Moody et al., 2011).

Alternative hosts are often blamed for confounding the reported sea lamprey marking rates on Great Lakes lake trout. This study confirms that suspicion in Lake Ontario, where the relative abundance of Chinook salmon affects the marking rate on lake trout. When we varied the proportion of hosts that were lake trout over the range observed (0.252 to 0.516) while holding the total abundance of hosts and the total number of marks on them
constant, the resulting marking rate on lake trout ranged as much as 0.034 (Figure 6). Here,
the lake trout marking rates are near their maximum when lake trout comprise at least
37% of the host abundance, but are reduced when the proportion of hosts that were lake
trout declines below the Model 1 inflection point of 32.6% (in 2006-2008 and 2014).

390 We hope that the evidence from switching demonstrated in this study and the methods 391 described will motivate further investigations into the host preference of sea lampreys. The 392 effect of alternative hosts on lake trout marking rates is also suspected in other Great 393 Lakes: siscowet lake trout in Lake Superior (Sitar et al., 2008), Chinook salmon in Lake 394 Michigan, and Chinook salmon and lake whitefish in Lake Huron (Hansen et al., 2016). One 395 impediment to such investigations is the lack of lake-wide absolute abundance estimates 396 for the alternative hosts, which are necessary to calculate the proportional abundance 397 relative to other hosts. Understanding the underlying causes of host switching could also be 398 investigated via more complex models that incorporate spatial overlap and quantify host 399 profitability to sea lampreys.

400 Acknowledgements

We thank the US Geological Survey - Great Lakes Science Center and the Great Lakes
Fishery Commission for financial support; the state, provincial, tribal, and federal agencies
for collecting annual information on Great Lakes lake trout and sea lampreys; the New York
State Department of Environmental Conservation, the Ontario Ministry of Natural

405 Resources, the USGS Great Lakes Science Center, and Cornell University for their data and 406 input on the Chinook Salmon estimates; Ted Treska for managing the GLFC Sea Lamprey 407 Marking Database; and Brian Lantry, Nick Johnson, and three anonymous reviewers for 408 suggestions that improved the manuscript. This research did not receive any specific grant 409 from funding agencies in the public, commercial, or not-for-profit sectors. Any use of trade, 410 firm, or product names is for descriptive purposes only and does not imply endorsement by 411 the U.S. Government. This research and manuscript were a component of the senior 412 author's PhD dissertation at Michigan State University. This is contribution number 20xx-413 **xx** of the Quantitative Fisheries Center at Michigan State University.

414 **References**

- Adams, J.V., Jones, M.L., Bence, J.R., *This issue*. What do annual marking rates really tell us
 about sea lampreys and lake trout? J. Great Lakes Res.
- 417 Adams. J.V., Jones. M.L., *In preparation*. Investigating misalignment of predator-prey

418 dynamics: Great Lakes lake trout and sea lamprey. Fisheries Research.

- Ahrens, R.N.M., Walters, C.J., Christensen, V., 2012. Foraging arena theory. Fish Fish. 13, 41–
 59.
- Andersson, M., Erlinge, S., 1977. Influence of predation on rodent populations. Oikos 29,
 591–597.
- 423 Angelstam, P., Lindström, E., Widén, P., 1984. Role of predation in short-term population
- fluctuations of some birds and mammals in Fennoscandia. Oecologia 62, 199–208.
- 425 Barber, J. M., et al. *This issue*. Measures for gauging program success. J. Great Lakes Res.

- Bayliss, D.E., 1982. Switching by *Lepsiella vinosa* (Gastropoda) in South Australian
 mangroves. Oecologia 54, 212–226.
- 428 Beamish, F.W.H., 1974. Swimming performance of adult sea lamprey, *Petromyzon marinus*,
- 429 in relation to weight and temperature. Trans. Am. Fish. Soc. 103, 355–358.
- 430 Begg, C.M., Begg, K.S., Du Toit, J.T., Mills, M.G.L., 2003. Sexual and seasonal variation in the
- 431 diet and foraging behaviour of a sexually dimorphic carnivore, the honey badger
- 432 (*Mellivora capensis*). J. Zool. 260, 301–316.
- 433 Bence, J.R., Bergstedt, R.A., Christie, G.C., Cochran, P.A., Ebener, M.P., Koonce, J.F., Rutter,
- 434 M.A., Swink, W.D., 2003. Sea lamprey (*Petromyzon marinus*) parasite-host interactions
- 435 in the Great Lakes. J. Great Lakes Res. 29, 253–282.
- 436 Bergstedt, R.A., Argyle, R.L., Krueger, C.C., Taylor, W.W., 2012. Bathythermal habitat use by
- 437 strains of Great Lakes- and Finger Lakes-origin lake trout in Lake Huron after a change
- in prey fish abundance and composition. Trans. Am. Fish. Soc. 141, 263–274.
- 439 Bergstedt, R.A., Schneider, C.P., 1988. Assessment of sea lamprey (*Petromyzon marinus*)
- 440 predation by recovery of dead lake trout (*Salvelinus namaycush*) from Lake Ontario,
- 441 1982-85. Can. J. Fish. Aquat. Sci. 45, 1406–1410.
- Bishop, D.L., Prindle, S.E., Verdoliva, F.J., 2019. 2018 Salmon River wild young-of-year
- 443 Chinook salmon seining program. In: Sect. 8 NYSDEC 2018 Annu. Report, Bur. Fish. Lake
- 444 Ontario Unit St. Lawrence River Unit to Gt. Lake Fish. Comm. Lake Ontario Comm. New
- 445 York State Department of Environmental Conservation, Albany, NY.
- 446 Brenden, T.O., Bence, J.R., Lantry, B.F., Lantry, J.R., Schaner, T., 2011. Population dynamics of
- 447 Lake Ontario lake trout during 1985–2007. North Am. J. Fish. Manag. 31, 962–979.

- 448 Cline, T.J., Kitchell, J.F., Bennington, V., McKinley, G.A., Moody, E.K., Weidel, B.C., 2014.
- 449 Climate impacts on landlocked sea lamprey: implications for host-parasite interactions
- 450 and invasive species management. Ecosphere 5, 68.
- 451 Coble, D.W., 1967. The white sucker population of South Bay, Lake Huron, and effects of the
- 452 sea lamprey on it. J. Fish. Res. Board Canada 24, 2117–2136.
- 453 Connerton, M.J., 2019. New York Lake Ontario and upper St. Lawrence River stocking
- 454 program 2018. In: Sect. 1 NYSDEC 2018 Annu. Report, Bur. Fish. Lake Ontario Unit St.
- 455 Lawrence River Unit to Gt. Lake Fish. Comm. Lake Ontario Comm. New York State
- 456 Department of Environmental Conservation, Albany, NY.
- 457 Connerton, M.J., Eckert, T.H., 2019. 2018 Lake Ontario fishing boat survey. In: Sect. 2
- 458 NYSDEC 2018 Annu. Report, Bur. Fish. Lake Ontario Unit St. Lawrence River Unit to Gt.
- 459 Lake Fish. Comm. Lake Ontario Comm. New York State Department of Environmental

460 Conservation, Albany, NY.

- 461 Cornell, H., 1976. Search strategies and the adaptive significance of switching in some
 462 general predators. Am. Nat. 110, 317–320.
- 463 Ebener, M.P., Bence, J.R., Bergstedt, R.A., Mullett, K.M., 2003. Classifying sea lamprey marks
- 464 on Great Lakes lake trout: observer agreement, evidence on healing times between
- 465 classes and recommendations for reporting of marking statistics. J. Great Lakes Res. 29,
 466 283–296.
- 467 Ebener, M.P., Brenden, T.O., Jones, M.L., 2010. Estimates of fishing and natural mortality
- 468 rates for four lake whitefish stocks in Northern Lakes Huron and Michigan. J. Great
- 469 Lakes Res. 36, 110–120.

- Fall, J., Fiksen, Ø., 2020. No room for dessert: a mechanistic model of prey selection in gutlimited predatory fish. Fish Fish. 21, 63–79.
- 472 Farmer, G.J., Beamish, F.W.H., 1973. Sea lamprey (*Petromyzon marinus*) predation on
- 473 freshwater teleosts. J. Fish. Res. Board Canada 30, 601–605.
- 474 Farmer, G.J., Beamish, F.W.H., Lett, P.F., 1977. Influence of water temperature on the growth
- 475 rate of the landlocked sea lamprey (*Petromyzon marinus*) and the associated rate of
- 476 host mortality. J. Fish. Res. Board Canada 34, 1373–1378.
- 477 Firkus, T., Murphy, C., Adams, J. V., Treska, T. J. *This issue*. Assessment of sea lamprey
- 478 wounds implications of healing progression on wound classification consistency and
- 479 accuracy. Journal of Great Lakes Research.
- 480 Fleischer, G.W., 1993. Status of coregonine fishes in the Laurentian Great Lakes. In: Todd,
- 481 T.N., Luczynski, M. (Eds.), Pol. Arch. Hydrobiol. Proceedings of the International

482 Symposium on Biology; Management of Coregonid Fishes, pp. 1–14.

- Fujii, K., Holling, C.S., Mace, P.M., 1986. A simple generalized model of attack by predators
 and parasites. Ecol. Res. 1, 141–156.
- 485 Hall-Scharf, B.J., Stallings, C.D., 2014. Experimental test of preference by a generalist
- 486 piscivore on morphologically- and behaviorally- different prey. J. Exp. Mar. Bio. Ecol.

487 460, 193–196.

- Hansen, M.J., Madenjian, C.P., Slade, J.W., Steeves, T.B., Almeida, P.R., Quintella, B.R., 2016.
- 489 Population ecology of the sea lamprey (*Petromyzon marinus*) as an invasive species in
- 490 the Laurentian Great Lakes and an imperiled species in Europe. Rev. Fish Biol. Fish. 26,
- 491 509–535.

492	Happel, A., Pattridge, R., Walsh, M., Rinchard, J., 2016. Assessing diet compositions of Lake
493	Ontario predators using fatty acid profiles of prey fishes. J. Great Lakes Res.
494	Happel, A., Rinchard, J., Czesny, S., 2017. Variability in sea lamprey fatty acid profiles
495	indicates a range of host species utilization in Lake Michigan. J. Great Lakes Res. 43,
496	182–188.
497	Harvey, C.J., Ebener, M.P., White, C.K., 2008. Spatial and ontogenetic variability of sea
498	lamprey diets in Lake Superior. J. Great Lakes Res. 34, 434–449.
499	Holbrook, C.M., Jubar, A.K., Barber, J.M., Tallon, K., Hondorp, D.W., 2016. Telemetry narrows
500	the search for sea lamprey spawning locations in the St. Clair-Detroit River System. J.
501	Great Lakes Res.
502	Hoyle, J.A., 2018. Lake Ontario and Bay of Quinte fish community index gill netting. In: Sect.
503	1.2 Lake Ontario Fish Communities Fish. 2017 Annu. Rep. Lake Ontario Manag. Unit.
504	Ontario Ministry of Natural Resources; Forestry, Picton, Ontario, Canada.
505	Hughes, R.N., 1979. Optimal diets under the energy maximization premise: the effects of
506	recognition time and learning. Am. Nat. 113, 209–221.
507	Joern, A., 1988. Foraging behavior and switching by the grasshopper sparrow Ammodramus
508	savannarum searching for multiple prey in a heterogeneous environment. Am. Midl.
509	Nat. 119, 225–234.
510	Johnson, J.E., Wright, G.M., Reid, D.M., Bowen II, C.A., Payne, N.R., 1995. Status of the cold-
511	water fish community in 1992. In: Ebener, M.P. (Ed.), State Lake Huron 1992. Great

512 Lakes Fishery Commission, Ann Arbor, MI, pp. 21–71.

- 513 King Jr., E.L., 1980. Classification of sea lamprey (*Petromyzon marinus*) attack marks on
- 514 Great Lakes lake trout (*Salvelinus namaycush*). Can. J. Fish. Aquat. Sci. 37, 1989–2006.
- 515 Kitchell, J.F., Breck, J.E., 1980. Bioenergetics model and foraging hypothesis for sea lamprey
- 516 (*Petromyzon marinus*). Can. J. Fish. Aquat. Sci. 37, 2159–2168.
- 517 Koonce, J.F., 1987. Application of models of lake trout/sea lamprey interaction to the
- 518 implementation of integrated pest management of sea lamprey in Lake Ontario. Great
- 519 Lakes Fishery Commission Project Completion Report, Ann Arbor, MI.
- 520 Krebs, J.R., Davies, N.B., 1987. An Introduction to Behavioural Ecology, 2nd ed. Sinauer
- 521 Associates, Sunderland, MA.
- 522 Lake, C., 2017. Chinook salmon net pen imprinting project. In: Sect. 7.1 Lake Ontario Fish
- 523 Communities Fish. 2016 Annu. Rep. Lake Ontario Manag. Unit. Ontario Ministry of
- 524 Natural Resources; Forestry, Picton, Ontario, Canada.
- 525 Lança, M.J., Machado, M., Ferreira, R., Alves-Pereira, I., Ruivo Quintella, B., Raposo de
- Almeida, P., 2013. Feeding strategy assessment through fatty acid profiles in muscles of
- adult sea lampreys from the western Iberian coast. Sci. Mar. 77, 281–291.
- 528 Lantry, B.F., Adams, J.V., Christie, G.C., Schaner, T., Bowlby, J., Keir, M., Lantry, J., Sullivan, P.,
- 529 Bishop, D., Treska, T., Morrison, B., 2015. Sea lamprey mark type, marking rate, and
- 530 parasite-host relationships for lake trout and other species in Lake Ontario. J. Great
- 531 Lakes Res. 41, 266–279.
- Lantry, B.F., Lantry, J.R., Connerton, M.J., 2018. Lake trout rehabilitation in Lake Ontario,
- 533 2017. In: Sect. 5 NYSDEC 2017 Annu. Report, Bur. Fish. Lake Ontario Unit St. Lawrence

- River Unit to Gt. Lake Fish. Comm. Lake Ontario Comm. New York State Department of
 Environmental Conservation, Albany, NY.
- 536 Lantry, J.R., Eckert, T.H., 2018. 2017 Lake Ontario fishing boat survey. In: Sect. 2 NYSDEC
- 537 2017 Annu. Report, Bur. Fish. Lake Ontario Unit St. Lawrence River Unit to Gt. Lake Fish.
- 538 Comm. Lake Ontario Comm. New York State Department of Environmental
- 539 Conservation, Albany, NY.
- Lantry, J.R., Schaner, T., Copeland, T., 2014. A Management Strategy for the Restoration of
- 541 Lake Trout in Lake Ontario, 2014 Update. Great Lakes Fishery Commission, Ann Arbor,
- 542 MI.
- Larson, G.L., Christie, G.C., Johnson, D.A., Koonce, J.F., Mullett, K.M., Sullivan, W.P., 2003. The
- history of sea lamprey control in Lake Ontario and updated estimates of suppression
 targets. J. Great Lakes Res. 29, 637–654.
- 546 Maitland, P.S., Renaud, C.B., Quintella, B.R., Close, D.A., Docker, M.F., 2015. Conservation of
- 547 Native Lampreys. In: Docker, M.F. (Ed.), Lampreys Biol. Conserv. Control. Vol. 1.
- 548 Springer, Dor, pp. 375–428.
- 549 Mason, D.M., Goyke, A., Brandt, S.B., 1995. A spatially explicit bioenergetics measure of
- habitat quality for adult salmonines: comparison between Lakes Michigan and Ontario.
- 551 May, R.M., 1977. Predators that switch. Nature 269, 103–104.
- 552 McLeod, D.V., Cottrill, R.A., Morbey, Y.E., 2011. Sea lamprey wounding in Canadian waters of
- Lake Huron from 2000 to 2009: temporal changes differ among regions. J. Great Lakes
- 554 Res. 37, 601–608.

- Merilaita, S., Ruxton, G.D., 2009. Optimal apostatic selection: how should predators adjust
 to variation in prey frequencies? Anim. Behav. 77, 239–245.
- 557 Moody, E.K., Weidel, B.C., Ahrenstorff, T.D., Mattes, W.P., Kitchell, J.F., 2011. Evaluating the
- growth potential of sea lampreys (*Petromyzon marinus*) feeding on siscowet lake trout
- 559 (*Salvelinus namaycush*) in Lake Superior. J. Great Lakes Res. 37, 343–348.
- 560 Mullett, K.M., Heinrich, J.W., Adams, J.V., Young, R.J., Henson, M.P., McDonald, R.B., Fodale,
- 561 M.F., 2003. Estimating lake-wide abundance of spawning-phase sea lampreys
- 562 (*Petromyzon marinus*) in the Great Lakes: extrapolating from sampled streams using
- regression models. J. Great Lakes Res. 29, 240–252.
- 564 Mumby, J.A., Larocque, S.M., Johnson, T.B., Stewart, T.J., Fitzsimons, J.D., Weidel, B.C., Walsh,
- 565 M.G., Lantry, J.R., Yuille, M.J., Fisk, A.T., 2018. Diet and trophic niche space and overlap of
- 566 Lake Ontario salmonid species using stable isotopes and stomach contents. J. Great
- 567 Lakes Res. 44, 1383–1392.
- Murdoch, W.W., 1969. Switching in general predators: experiments on predator specificity
 and stability of prey populations. Ecol. Monogr. 39, 335–354.
- 570 Murdoch, W.W., Avery, S., Smyth, M.E.B., 1975. Switching in predatory fish. Ecology 56,
 571 1094–1105.
- 572 Neff, M.R., Bhavsar, S.P., Chin, J.X.Y., 2012. Spatial and temporal trends of muscle lipid
- 573 content in Great Lakes fishes: 1970s-2008. Can. J. Fish. Aquat. Sci. 69, 2007+.
- 574 Nowicki, S.M., 2008. Healing, classification and hematological assessments of sea lamprey
- 575 (*Petromyzon marinus*) wounds on lake trout (*Salvelinus namaycush*). (Master's thesis).
- 576 Northern Michigan University.

577	Patterson, B.R., Benjamin, L.K., Messier, F., 1998. Prey switching and feeding habits of
578	eastern coyotes in relation to snowshoe hare and white-tailed deer densities. Can. J.
579	Zool. 76, 1885–1897.

- 580 Prichard, C.G., Bence, J.R., 2013. Estimating wounding of lake trout by sea lamprey in the
- 581 upper Great Lakes: allowing for changing size-specific patterns. J. Great Lakes Res. 39,

582 110–119.

583 Prindle, S.E., Bishop, D.L., 2018. Population characteristics of pacific salmonines collected at

the Salmon River Hatchery 2017. In: Sect. 9 NYSDEC 2017 Annu. Report, Bur. Fish. Lake

585 Ontario Unit St. Lawrence River Unit to Gt. Lake Fish. Comm. Lake Ontario Comm. New

586 York State Department of Environmental Conservation, Albany, NY.

587 Prindle, S.E., Bishop, D.L., 2019. Population characteristics of Pacific salmonines collected at

the Salmon River Hatchery 2018. In: Sect. 9 NYSDEC 2018 Annu. Report, Bur. Fish. Lake

589 Ontario Unit St. Lawrence River Unit to Gt. Lake Fish. Comm. Lake Ontario Comm. New

590 York State Department of Environmental Conservation, Albany, NY.

591 Pycha, R.L., King, G.R., 1975. Changes in the lake trout population of southern Lake Superior

- in relation to the fishery, the sea lamprey, and stocking, 1950-70. Great Lakes Fishery
- 593 Commission Tech. Rep. 28, Ann Arbor, MI.

Rand, P.S., Stewart, D.J., 1998. Prey fish exploitation, salmonine production, and pelagic

- food web efficiency in Lake Ontario. Can. J. Fish. Aquat. Sci. 55, 318–327.
- 596 R Core Team, 2018. R: a language and environment for statistical computing. R Foundation
- 597 for Statistical Computing, Vienna, Austria.

- 598 Rubega, M., Inouye, C., 1994. Prey switching in red-necked phalaropes *Phalaropus lobatus*:
- 599 feeding limitations, the functional response and water management at Mono Lake,
- 600 California, USA. Biol. Conserv. 70, 205–210.
- 601 Rutter, M.A., Bence, J.R., 2003. An improved method to estimate sea lamprey wounding rate
- on hosts with application to lake trout in Lake Huron. J. Great Lakes Res. 29, 320–331.
- 603 Sanderson, M.J., Todd, M.T., Prindle, S.E., 2019. 2018 New York cooperative trout and
- 604 salmon pen-rearing projects. In: Sect. 10 NYSDEC 2018 Annu. Report, Bur. Fish. Lake
- 605 Ontario Unit St. Lawrence River Unit to Gt. Lake Fish. Comm. Lake Ontario Comm. New
- 606 York State Department of Environmental Conservation, Albany, NY.
- 607 Schneider, C.P., Owens, R.W., Bergstedt, R.A., O'Gorman, R., 1996. Predation by sea lamprey
- 608 (*Petromyzon marinus*) on lake trout (*Salvelinus namaycush*) in southern Lake Ontario,
- 609 1982-1992. Can. J. Fish. Aquat. Sci. 53, 1921–1932.
- 610 Schoener, T.W., 1971. Theory of feeding strategies. Annu. Rev. Ecol. Syst. 2, 369–404.
- 611 Seber, G.A.F., 1970. The effects of trap response on tag recapture estimates. Biometrics 26,
 612 13–22.
- 613 Silva, S., Araújo, M.J., Bao, M., Mucientes, G., Cobo, F., 2014. The haematophagous feeding
- 614 stage of anadromous populations of sea lamprey Petromyzon marinus: low host
- 615 selectivity and wide range of habitats. Hydrobiologia 734, 187–199.
- 616 Sitar, S.P., Morales, H.M., Mata, M.T., Bastar, B.B., Dupras, D.M., Kleaver, G.D., Rathbun, K.D.,
- 617 2008. Survey of siscowet lake trout at their maximum depth in Lake Superior. J. Great
- 618 Lakes Res. 34, 276–286.

619	Smith, B.R., Tibbles, J., 1980. Sea lamprey (<i>Petromyzon marinus</i>) in Lakes Huron, Michigan,
620	and Superior: history of invasion and control, 1936-78. Can. J. Fish. Aquat. Sci. 37, 1780–
621	1801.
622	Spangler, G.R., Robson, D.S., Regier, H.A., 1980. Estimates of lamprey-induced mortality in
623	whitefish, Coregonus clupeaformis. Can. J. Fish. Aquat. Sci. 37, 2146–2150.
624	Stapanian, M.A., Madenjian, C.P., Witzel, L.D., 2006. Evidence that sea lamprey control led to
625	recovery of the burbot population in Lake Erie. Trans. Am. Fish. Soc. 135, 1033–1043.
626	Stewart, D.J., Ibarra, M., 1991. Predation and production by salmonine fishes in Lake
627	Michigan, 1978–88. Can. J. Fish. Aquat. Sci. 48, 909–922.
628	Stewart, T.J., Bence, J.R., Bergstedt, R.A., Ebener, M.P., Lupi, F., Rutter, M.A., 2003.
629	Recommendations for assessing sea lamprey damages: toward optimizing the control
630	program in the Great Lakes. J. Great Lakes Res. 29, 783–793.
631	Swink, W.D., 2003. Host selection and lethality of attacks by sea lampreys (Petromyzon
632	<i>marinus</i>) in laboratory studies. J. Great Lakes Res. 29, 307–319.
633	Taylor, R.J., 1977. The value of clumping to prey: experiments with a mammalian predator.
634	Oecologia 30, 285–294.
635	Van Baalen, M., Křivan, V., Van Rijn, P.C.J., Sabelis, M.W., 2001. Alternative food, switching
636	predators, and the persistence of predator-prey systems. Am. Nat. 157, 512.
637	Van Leeuwen, E., Brannstrom, A., Jansen, V.A.A., Dieckmann, U., Rossberg, A.G., 2013. A
638	generalized functional response for predators that switch between multiple prey
639	species. J. Theor. Biol. 328, 89–98.

- 640 Walters, C.J., 1986. Adaptive Management of Renewable Resources. Macmillan, New York,641 NY.
- 642 Weidel, B.C., Connerton, M.J., Holden, J.P., 2018. Bottom trawl assessment of Lake Ontario
- 643 prey fishes. In: Sect. 12 NYSDEC 2017 Annu. Report, Bur. Fish. Lake Ontario Unit St.
- 644 Lawrence River Unit to Gt. Lake Fish. Comm. Lake Ontario Comm. New York State
- 645 Department of Environmental Conservation, Albany, NY.
- 646 Weitkamp, L.A., Weitkamp, L.A., Hinton, S.A., Bentley, P.J., 2015. Seasonal abundance, size,
- 647 and host selection of western river (*Lampetra ayresii*) and Pacific (*Entosphenus*
- *tridentatus*) lampreys in the Columbia River estuary. Fish. Bull. 113, 213–226.
- 649 Young, R.J., Christie, G.C., McDonald, R.B., Cuddy, D.W., Morse, T.J., Payne, N.R., 1996. Effects
- 650 of habitat change in the St. Marys River and northern Lake Huron on sea lamprey
- 651 (*Petromyzon marinus*) populations. Can. J. Fish. Aquat. Sci. 53, 99–104.
- 652 Yuille, M.J., Holden, J.P., 2017. Chinook salmon. In: Sect. 6.2 Lake Ontario Fish Communities
- Fish. 2016 Annu. Rep. Lake Ontario Manag. Unit. Ontario Ministry of Natural Resources;
- 654 Forestry, Picton, Ontario, Canada.

656 Tables

657 Table 1 . Lake trout relative abundance	(expressed as the proportion of lake trout and
--	--

658 Chinook salmon hosts) and lake trout and Chinook salmon parasitism rates (N=number of

659	hosts observed, A1=number of A1 or "fresh	" sea lamprey-induced	marks per host).
	,	1 2	1 2

	Lake trout	Lake trout	Lake trout	Chinook salmon	Chinook salmon
Year	relative abundance	Ν	A1	Ν	A1
2000	0.516	993	0.023	646	0.000
2001	0.466	991	0.023	657	0.000
2002	0.391	909	0.008	624	0.002
2003	0.387	943	0.020	923	0.015
2004	0.398	891	0.018	744	0.009
2005	0.354	468	0.032	753	0.012
2006	0.316	609	0.025	630	0.006
2007	0.271	399	0.030	481	0.035
2008	0.252	554	0.016	669	0.016
2009	0.370	613	0.010	897	0.012
2010	0.414	785	0.015	719	0.000
2011	0.407	926	0.004	625	0.003
2012	0.402	871	0.025	574	0.003
2013	0.328	1122	0.017	584	0.003
2014	0.284	1268	0.010	675	0.006

- **Table 2**. Estimated parameters relating host length to sea lamprey-induced marking rate
- 662 for both species combined and for lake trout and Chinook salmon separately.

Species	Alpha	Beta	Theta
Both	0.0200	587	0.0150
Lake trout	0.0207	383	0.0161
Chinook salmon	0.0197	718	0.0105

664 **Figure captions**

Figure 1. Annual length frequency distributions of lake trout and Chinook salmon capturedin Lake Ontario 2000-2014. Each line represents a separate year.

667 Figure 2. Relation between the proportion of hosts that were lake trout and the proportion 668 of sea lamprey-induced marks that were on lake trout in Lake Ontario, assuming lake trout 669 and Chinook salmon were the only available hosts. Symbols represent annual observations 670 (2000-2014), symbol size represents number of hosts examined for marks, and lines 671 represent predictions from Model 1 with 95% confidence intervals. The diagonal dotted 672 line represents the null model, with no preference and no switching. The two triangles 673 denote the switch points below which sea lampreys prefer Chinook salmon (left) and above 674 which sea lampreys prefer lake trout (right).

Figure 3. Sea lamprey-induced marking rate as a function of host size for just lake trout,
just Chinook salmon, and both species combined in Lake Ontario 2000-2014. Symbols
represent observed proportions for each 50 mm size group, symbol size represents the
sample size, and lines represent predictions from the negative binomial model.

Figure 4. Proportion of sea lamprey-induced marks that were on lake trout (relative to
Chinook salmon) over time in Lake Ontario 2000-2014. Points represent observations,
lines represent predictions from three models based on (1) the proportion of hosts that
were lake trout (solid), (2) the proportion of hosts that were lake trout and the size of the

mixed hosts (dashed), and (3) the proportion of hosts that were lake trout and speciesspecific host size (dotted).

Figure 5. Relation between lake trout abundance and the proportion of sea lampreyinduced marks that were on lake trout (relative to Chinook salmon) in Lake Ontario.
Symbols represent annual observations (2000-2014), symbol size represents number of
hosts captured, and lines represent predictions from Model 1 with 95% confidence
intervals.

Figure 6. Model 1 predicted range of sea lamprey-induced marking rates on lake trout in
Lake Ontario (shading) when the proportion of hosts that were lake trout (relative to
Chinook salmon) varies over the range observed (0.252 to 0.516) while holding the total
abundance of hosts and the total number of marks on them constant. Points represent
observed marking rates, line represents predicted marking rates from Model 1 for
observed proportional abundances.





Figure 3 Click here to download high resolution image



Ν





